

# Hg<sup>2+</sup>-induced leakage of electrolytes and inhibition of NO<sub>3</sub><sup>-</sup> utilization in *Nostoc calcicola*

## Role of interacting cations

Chandra Bhushan Singh and S. P. Singh

Algal Research Laboratory, Centre of Advanced Study in Botany, Banaras Hindu University, Varanasi-221005, India

Received July 19, 1990

**Summary.** The effect of mercury (Hg<sup>2+</sup>) in the absence and presence of methylmercury (CH<sub>3</sub>Hg<sup>+</sup>), cadmium (Cd<sup>2+</sup>), copper (Cu<sup>2+</sup>), nickel (Ni<sup>2+</sup>) and calcium (Ca<sup>2+</sup>) on *Nostoc calcicola* Bréb. has been studied in terms of electrolyte leakage, NO<sub>3</sub><sup>-</sup> uptake and in vivo nitrate reductase (NR) activity to discover any possible correlation among such parameters under Hg<sup>2+</sup> stress. Leakage of electrolytes from Hg<sup>2+</sup>-treated cyanobacterial cells was directly proportional to Hg<sup>2+</sup> concentrations and exposure time. In comparison to NO<sub>3</sub><sup>-</sup> uptake, an about 60-fold slower rate of NR activity was observed in the untreated cultures, the former being five times more Hg<sup>2+</sup>-sensitive. A non-competitive synergistic interaction of Hg<sup>2+</sup> with CH<sub>3</sub>Hg<sup>+</sup> or Cd<sup>2+</sup> and antagonistic with that of Ni<sup>2+</sup> or Ca<sup>2+</sup> has been observed for both the processes of NO<sub>3</sub><sup>-</sup> utilization. The antagonistic interaction of Cu<sup>2+</sup> with Hg<sup>2+</sup> in terms of NO<sub>3</sub><sup>-</sup> uptake and synergistic with respect to NR activity, has been attributed to the dual bonding preference of Cu<sup>2+</sup> for cellular ligands. These findings suggest that (a) a statistically significant correlation exists among such parameters; (b) Hg<sup>2+</sup> predominantly attacks the cyanobacterial cell membrane; (c) Hg<sup>2+</sup> inhibits NO<sub>3</sub><sup>-</sup> utilization; (d) the presence of other cations increases or decreases the inhibitory actions of Hg<sup>2+</sup>.

**Key words:** Hg<sup>2+</sup> toxicity – *Nostoc calcicola* – Electrolyte leakage – NO<sub>3</sub><sup>-</sup> uptake – Nitrate reductase – Metal interactions

## Introduction

The extremely toxic metal cation of mercury (Hg<sup>2+</sup>) has no known biological requirement. In natural habi-

tats, the heavy metal may originate from domestic sewage, agricultural run-off and industrial or pharmaceutical wastes and ultimately enters biological systems as the active cation to interact with intracellular targets (Blinn et al. 1977; Sigmon et al. 1977; Jeffries 1982).

Hg<sup>2+</sup> increases cell membrane permeability and induces the leakage of intracellular ions in eukaryotic algae (Shieh and Barber 1973; Rai et al. 1981); however, there are no similar reports on prokaryotic cyanobacteria. The transmembrane uptake of nitrate (NO<sub>3</sub><sup>-</sup>) is largely an energy-dependent process and has been reported to be quite sensitive to Cd<sup>2+</sup> in *Anacystis nidulans* (Singh and Yadava 1983) and to Cr<sup>2+</sup> or Sn<sup>2+</sup> in *Anabaena doliolum* (Rai and Dubey 1989). Nitrate reductase (NR), the key enzyme for NO<sub>3</sub><sup>-</sup> reduction, operates at the expense of electrons derived through photosynthesis (Guerrero and Lara 1987), a process which is extremely sensitive to Hg<sup>2+</sup> (Stratton et al. 1979; Singh and Singh 1987; Singh et al. 1989). Recently, in vivo NR activity was found to be quite sensitive to Cr<sup>2+</sup> and Sn<sup>2+</sup> (Rai and Dubey 1989). However, the reports so far available provide no information on the biology of Hg<sup>2+</sup> toxicity to cyanobacteria in terms of electrolyte leakage, concurrently with NO<sub>3</sub><sup>-</sup> uptake and in vivo NR activity.

Apart from various biotic factors, Hg<sup>2+</sup> toxicity is greatly influenced by the presence of other interacting cations (Stratton and Corke 1979; Singh and Singh 1987; Singh et al. 1987). Since the natural habitats are generally characterized by the co-existence of a large number of toxic and non-toxic cations (Whitton 1970; Henriksen and Wright 1978; Strarodub et al. 1987) and since algal growth media also contain a large number of cations, it is necessary to study Hg<sup>2+</sup> toxicity in the presence of other interacting cations.

The present study, therefore, was aimed at investigating such aspects in order to discover a possible correlation in electrolyte leakage, NO<sub>3</sub><sup>-</sup> uptake and in vivo NR activity in the diazotrophic cyanobacterium *Nostoc calcicola* Bréb. exposed to either Hg<sup>2+</sup> alone or to its combinations with CH<sub>3</sub>Hg<sup>+</sup>, Cd<sup>2+</sup>, Cu<sup>2+</sup>, Ni<sup>2+</sup> or Ca<sup>2+</sup>.

## Materials and methods

**Experimental organism and growth conditions.** *Nostoc calcicola* Bréb. was axenically grown in liquid growth medium (Allen and Arnon 1955) free from any combined nitrogen sources. The cultures were routinely maintained in 200 ml medium, contained in 500-ml conical flasks and illuminated with cool fluorescent light ( $50 \mu\text{mol photon} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$ ) in a culture room at  $24 \pm 1^\circ \text{C}$  with an 18-h/6-h light/dark regime. Protein was quantified by the method of Lowry et al. (1951), modified by Herbert et al. (1971), using lysozyme (Sigma, USA) as standard.

**Electrolyte leakage.** Exponentially growing *N. calcicola* cells were harvested by centrifugation and, after washing three times with sterile distilled water, inoculated into fresh sterile liquid medium (500  $\mu\text{g}$  protein/ml culture), containing graded concentrations of  $\text{Hg}^{2+}$  (0.5–20.0  $\mu\text{M}$  as  $\text{HgCl}_2$ ; BDH, UK). Such sets were phototrophically incubated for 16 h and the samples (25.0 ml), withdrawn at 2-h intervals, were centrifuged and the pellets resuspended in sterile distilled water for 30 min to ensure the leakage of electrolytes. Such solutions were further centrifuged and the ion content of the resultant cell-free supernatant was monitored by recording the specific conductance in a conductivity meter (sensitivity range 0.02–200 mS/mg protein; Systronics, India). The values, expressed as  $\mu\text{S}/\text{mg}$  protein, were obtained after deducting the background specific conductance of sterile distilled water.

**Nitrate uptake.** The  $\text{N}_2$ -grown log-phase *N. calcicola* cells were inoculated in a sterile basal medium (500  $\mu\text{g}$  protein/ml culture), containing graded concentrations of  $\text{NO}_3^-$  (1.0–25.0 mM; BDH, UK), to obtain the saturating  $\text{NO}_3^-$  concentrations for optimum uptake. To assess the 50% inhibitory  $\text{Hg}^{2+}$  concentration, the cyanobacterial cells were exposed to graded concentrations of  $\text{Hg}^{2+}$  (1.25–20.0  $\mu\text{M}$ ) at saturating  $\text{NO}_3^-$  concentration (20.0 mM) in the basal medium. This inhibitory  $\text{Hg}^{2+}$  concentration (2.5  $\mu\text{M}$ ) was used in subsequent experiments on metal interactions, involving equimolar concentrations (2.5  $\mu\text{M}$  each) of either  $\text{CH}_3\text{Hg}^+$  (as  $\text{CH}_3\text{HgCl}$ ; Wilson Lab., India),  $\text{Cd}^{2+}$ ,  $\text{Cu}^{2+}$ ,  $\text{Ni}^{2+}$  or  $\text{Ca}^{2+}$  (as  $\text{CdCl}_2$ ,  $\text{CuCl}_2$ ,  $\text{NiCl}_2$  or  $\text{CaCl}_2$ , respectively; BDH, UK) and graded  $\text{NO}_3^-$  concentrations (1.25–20.0 mM). After treatments, the sets were incubated phototrophically for 6 h and the amount of  $\text{NO}_3^-$  taken up by the cyanobacterial cells was determined by estimating the disappearance of  $\text{NO}_3^-$  from the external medium. The aliquots were withdrawn at regular 1-h intervals and  $\text{NO}_3^-$  in the supernatant fluid analyzed using the Brucine/ $\text{H}_2\text{SO}_4$  method (Nicholas and Nason 1957).

**Nitrate reductase activity.** The in vivo NR activity in  $\text{Hg}^{2+}$ -treated (2.5–20.0 mM) *N. calcicola* cells was monitored under identical experimental conditions to  $\text{NO}_3^-$  uptake (see above) by estimating the amount of  $\text{NO}_2^-$  produced in the  $\text{NO}_3^-$ -supplemented (20.0 mM) basal medium. Metal-interaction experiments were also performed in the same way as for  $\text{NO}_3^-$  uptake, except that 12.5  $\mu\text{M}$  concentration (the 50% inhibitory  $\text{Hg}^{2+}$  level) of each interacting cation was used as separate bimetallic combinations with  $\text{Hg}^{2+}$ . The samples (5.0 ml), withdrawn at regular 30-min intervals, were permeabilized with 0.5 ml toluene (Merck, India), followed by rigorous shaking and incubation at  $4^\circ \text{C}$  for 10 min. Such samples were centrifuged to remove the toluene layer and the cell extract subjected to colorimetric estimations by using the azo-coupling method of Snell and Snell (1949).

## Results

### Electrolyte leakage under $\text{Hg}^{2+}$ stress

The two-way analysis of variance (ANOVA) of the data represented in Table 1 shows that the leakage of elec-

**Table 1.** Leakage of electrolytes from *Nostoc calcicola* cells exposed to  $\text{Hg}^{2+}$

$\text{Hg}^{2+}$ ( $\mu\text{M}$ )	Leakage ( $\mu\text{S}/\text{mg}$ protein) after incubation for				
	2 h	4 h	8 h	12 h	16 h
(control)	4	7	10	13	18
0.5	16	26	36	44	52
1.25	20	44	82	102	121
2.5	30	64	118	156	182
5.0	52	90	142	182	208
10.0	64	105	172	208	216
15.0	66	108	178	210	218
20.0	72	118	182	214	222

The data for electrolyte leakage (Table 1) and in vivo NR activity in relation various metal treatments (Fig. 4) and exposure time, were verified by analysis of variance (ANOVA) to test their significance at a particular probability level, and the variance ratio ( $F$ ) calculated by the equation  $F = (\text{treatment mean square})/(\text{residual mean square})$ . The value for incubations was  $F_{4,28} = 28.864$ ,  $P < 0.005$ ; for  $\text{Hg}^{2+}$  concentrations  $F_{7,28} = 26.480$ ,  $P < 0.005$ .

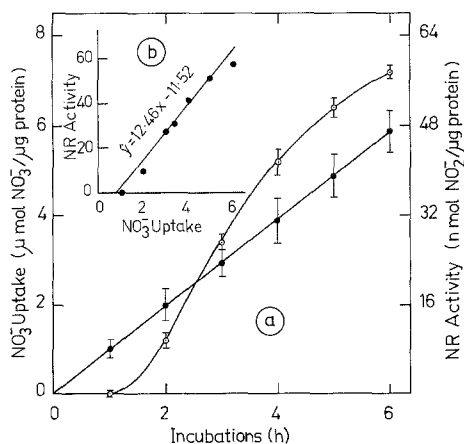
trolytes from  $\text{Hg}^{2+}$ -treated *N. calcicola* cells varied significantly ( $P < 0.005$ ) with respect to increasing metal concentrations and exposure time. Apart from cell death at  $\text{Hg}^{2+}$  concentrations exceeding 10.0  $\mu\text{M}$ , the leakage of electrolytes was also found operative for the untreated cyanobacterial cells at a much slower pace (0.75  $\mu\text{S} \cdot \text{mg protein}^{-1} \cdot \text{h}^{-1}$ ). A progressive increase in the extent of electrolyte leakage was observed for cells treated with 0.5–5.0  $\mu\text{M}$   $\text{Hg}^{2+}$ . However, on exposure of more than 16 h to 10.0  $\mu\text{M}$   $\text{Hg}^{2+}$  showed maximum rate of electrolyte efflux (i.e. 16.75  $\mu\text{S} \cdot \text{protein}^{-1} \cdot \text{h}^{-1}$ , as determined between 4–8-h exposures). Even higher  $\text{Hg}^{2+}$  concentrations (10.0–20.0  $\mu\text{M}$ ) neither stimulated the rate of electrolyte leakage nor caused a proportional increase in the specific conductance; instead, severe cell lysis was observed microscopically.

### $\text{NO}_3^-$ uptake and reduction in untreated cultures

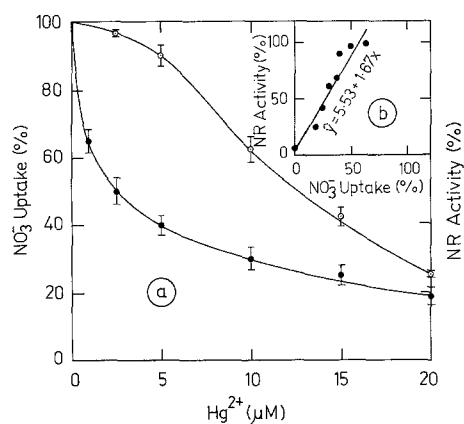
Compared to reduction of  $\text{NO}_3^-$  at cell interior (16.0 nmol  $\text{NO}_2^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$ ), its uptake from external medium proceeded  $>60$  times faster (1.0  $\mu\text{mol NO}_3^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$ ) with linearity over 6 h (Fig. 1a). The respective rates have been calculated from the linear portion of the curves (i.e. between 2–4 h of incubation). A distinct lag of 1 h was observed for the in vivo NR activity in untreated cyanobacterial cells. Statistical analyses of the data established a significant correlation ( $r = 0.988$ ,  $df 4$ ,  $P < 0.01$ ;  $\hat{y} = 12.4x - 11.52$ ) between the two processes in terms of  $\text{NO}_3^-$  utilization (Fig. 1b).

### $\text{Hg}^{2+}$ sensitivity of $\text{NO}_3^-$ uptake and in vivo NR activity

A comparison of the slopes in Fig. 2a and the corresponding 50% inhibitory  $\text{Hg}^{2+}$  concentrations (2.5  $\mu\text{M}$



**Fig. 1.** Pattern of NO<sub>3</sub><sup>-</sup> uptake and in vivo nitrate reductase activity in *N. calicicola* at 20.0 mM KNO<sub>3</sub>. (a) Direct plot of NO<sub>3</sub><sup>-</sup> uptake (●) versus NR activity (○). Values are mean  $\pm$  3 SE;  $r = 0.988$ ,  $df$  4,  $P < 0.01$ . (b) Regression analysis.  $x = \text{NO}_3^-$  uptake and  $\hat{y} =$  in vivo nitrate reductase activity

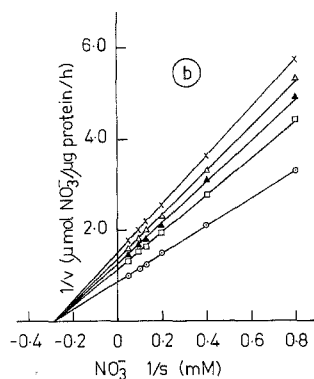


**Fig. 2.** A comparison of NO<sub>3</sub><sup>-</sup> uptake and in vivo nitrate reductase activity at 20.0 mM KNO<sub>3</sub> in *N. calicicola* after 6-h exposure to graded Hg<sup>2+</sup> concentrations. (a) Direct plot of NO<sub>3</sub><sup>-</sup> uptake (●) versus NR activity (○). Values are mean  $\pm$  3 SE;  $r = 0.916$ ,  $df$  4,  $P < 0.05$ . (b) Regression analysis.  $x = \text{NO}_3^-$  uptake and  $\hat{y} =$  in vivo nitrate reductase activity

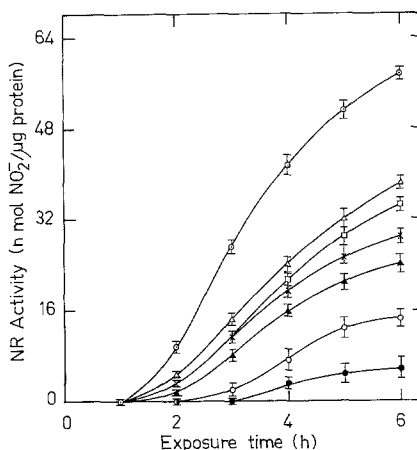
and 12.5 μM, respectively; as determined after 6-h exposures) suggested that the heavy metal inhibited NO<sub>3</sub><sup>-</sup> uptake more severely than its subsequent reduction at cell interior (about fivefold difference). While even low Hg<sup>2+</sup> concentrations (<5.0 μM) were most effective against the former, acute inhibition of the latter could be achieved only at more than 5.0 μM Hg<sup>2+</sup>. The data in Fig. 2b further substantiate a positive, significant correlation ( $r = 0.916$ ,  $df$  4,  $P < 0.05$ ) and regression ( $\hat{y} = 5.53 + 1.67x$ ) between these metabolic events in terms of Hg<sup>2+</sup> inhibition.

**Interaction of CH<sub>3</sub>Hg<sup>+</sup>, Cd<sup>2+</sup>, Cu<sup>2+</sup>, Ni<sup>2+</sup> or Ca<sup>2+</sup> with Hg<sup>2+</sup>**

**NO<sub>3</sub><sup>-</sup> uptake kinetics.** NO<sub>3</sub><sup>-</sup> uptake at graded KNO<sub>3</sub> concentrations (1.25–20.0 mM), followed Michaelis-



**Fig. 3.** Lineweaver-Burk plots for NO<sub>3</sub><sup>-</sup> uptake in *N. calicicola*. (a) At graded KNO<sub>3</sub> concentrations and 2.5 μM Hg (50% inhibitory concentration) with other metal cations (equimolar); metal-less control (○—○), Hg<sup>2+</sup> alone (x—x), Hg<sup>2+</sup> + Cd<sup>2+</sup> (○—○) and Hg<sup>2+</sup> + CH<sub>3</sub>Hg (●—●). (b) Showing interaction of Hg<sup>2+</sup> (2.5 μM) with other metal cations (equimolar): metal-less control (○—○), Hg<sup>2+</sup> alone (x—x), Hg<sup>2+</sup> + Ni<sup>2+</sup> (Δ—Δ), Hg<sup>2+</sup> + Cu<sup>2+</sup> (▲—▲) and Hg<sup>2+</sup> + Ca<sup>2+</sup> (□—□)



**Fig. 4.** In vivo nitrate reductase activity in *N. calicicola* at 20.0 mM KNO<sub>3</sub> showing interaction of Hg<sup>2+</sup> (50% inhibitory concentration, 12.5 μM) with other metal cations (equimolar): metal-less control (○—○), Hg<sup>2+</sup> alone (x—x), Hg<sup>2+</sup> + Cd<sup>2+</sup> (○—○), Hg<sup>2+</sup> + CH<sub>3</sub>Hg (●—●), Hg<sup>2+</sup> + Ni<sup>2+</sup> (Δ—Δ), Hg<sup>2+</sup> + Cu<sup>2+</sup> (▲—▲) and Hg<sup>2+</sup> + Ca<sup>2+</sup> (□—□); values are mean  $\pm$  3 SE;  $F_{\text{hours } 4,24} = 19.267$  and  $F_{\text{metal combinations } 6,24} = 20.437$ ,  $P < 0.005$

Menten kinetics with  $K_m$  and  $V_{\text{max}}$  values of 3.57 mM and 1.18 μmol NO<sub>3</sub><sup>-</sup> · μg protein<sup>-1</sup> · h<sup>-1</sup>, respectively (Fig. 3a, b). The simultaneous addition of 2.5 μM Hg<sup>2+</sup> (the 50% inhibitory level) did not alter the  $K_m$ ; nevertheless,  $V_{\text{max}}$  was reduced to 0.67 μmol NO<sub>3</sub><sup>-</sup> · μg protein<sup>-1</sup> · h<sup>-1</sup>. The bimetallic combinations of Hg<sup>2+</sup> with either Cd<sup>2+</sup> or CH<sub>3</sub>Hg<sup>+</sup>, further reduced the  $V_{\text{max}}$

to 0.61 and 0.49  $\mu\text{mol NO}_3^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$ , respectively (Fig. 3a). However,  $\text{Hg}^{2+}$  in combination with either  $\text{Ni}^{2+}$ ,  $\text{Cu}^{2+}$  or  $\text{Ca}^{2+}$  enhanced the corresponding  $V_{\text{max}}$  values (0.74, 0.80 and 0.91  $\mu\text{mol NO}_3^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$ ) over the cyanobacterial cells exposed to  $\text{Hg}^{2+}$  alone (Fig. 3b).

**In vivo nitrate reductase activity.** Interaction experiments on similar lines showed that the reduction of  $\text{NO}_3^-$  was reduced to half (8.0 nmol  $\text{NO}_2^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$ ) in presence of 12.5  $\mu\text{M Hg}^{2+}$ , but the simultaneous addition of equimolar concentrations (12.5  $\mu\text{M}$  each) of either  $\text{Ca}^{2+}$  or  $\text{Ni}^{2+}$  with  $\text{Hg}^{2+}$ , resulted in the recovery of enzyme activity (9.2 and 9.6 nmol  $\text{NO}_2^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$  respectively; Fig. 4). In contrast,  $\text{Hg}^{2+}$  in combination with either  $\text{Cu}^{2+}$ ,  $\text{Cd}^{2+}$  or  $\text{CH}_3\text{Hg}^+$  reduced the in vivo activity rates of nitrate reductase to 7.2, 3.2 and 1.6 nmol  $\text{NO}_2^- \cdot \mu\text{g protein}^{-1} \cdot \text{h}^{-1}$ , respectively. Such variations in the NR activity rates in response to different  $\text{Hg}^{2+}$  combinations and exposure time were found to be statistically significant at less than 0.5% probability level.

## Discussion

The negligible efflux of electrolytes from untreated *N. calcicola* cells (0.75  $\mu\text{S} \cdot \text{mg protein}^{-1} \cdot \text{h}^{-1}$ ; Table 1) suggests that the cyanobacterial cell membrane actively regulates the transmembrane movement of electrolytes. The proportional increase in electrolyte leakage corresponding to the exposure to less than 5.0  $\mu\text{M Hg}^{2+}$  suggests it arises as a consequence of the irreversible binding of  $\text{Hg}^{2+}$  with the adsorption sites on the cell surface, as reported for *Anabaena inaequalis* (Stratton et al. 1979). While the irregular enhancement in the electrolyte loss at higher  $\text{Hg}^{2+}$  concentrations (5.0–20.0  $\mu\text{M}$ ) may be attributed to the disruption of the cell membrane, as reported for ion loss in other cyanobacteria induced by Cu, Cd, Cr or Pb (Mierle and Stokes 1976; Singh and Yadava 1986; Rai and Raizada 1987, 1988). A corollary of these findings is that the cell membrane represents the primary site of  $\text{Hg}^{2+}$  action in *N. calcicola* and even very low  $\text{Hg}^{2+}$  concentrations may disrupt its integrity, leading to the uncontrolled leakage of intracellular electrolytes. The cell lysis at  $\text{Hg}^{2+}$  concentrations exceeding 10.0  $\mu\text{M}$  possibly represents the ultimate fate of the cell membrane. Thus, it is suggested that the electrolyte leakage may be taken as a reliable criterion to assess cell membrane integrity in cyanobacteria.

In comparison to  $\text{NO}_3^-$  uptake, the NR activity not only showed a lag of 1 h but also a slow activity rate in untreated *N. calcicola* cells (about 60-fold difference; Fig. 1). Such data are in conformity with those of other cyanobacteria in that  $\text{NO}_3^-$  enters the cell via a transport system and the anion itself acts as the inducer of nitrate reductase (Flores et al. 1983). The overall comparison of the inhibition events in general, and in particular of the corresponding 50% inhibitory  $\text{Hg}^{2+}$  concentrations for  $\text{NO}_3^-$  uptake (2.5  $\mu\text{M}$ ) and NR activity

(12.5  $\mu\text{M}$ ), suggests a five-fold enhanced  $\text{Hg}^{2+}$  sensitivity of the former (Fig. 2). The severe inhibition of enzyme activity exclusively at more than 5.0  $\mu\text{M Hg}^{2+}$ , and  $\text{NO}_3^-$  uptake at less than that, suggests that sub-optimal availability of  $\text{NO}_3^-$  is needed to induce the nitrate reductase in the cell interior. The unaltered  $K_m$  compared to  $V_{\text{max}}$  indicates a condition of non-competitive interaction between  $\text{NO}_3^-$  and  $\text{Hg}^{2+}$  and rules out the possibility of a common site of entry for both ions or a competitive interaction (Fig. 3). However, such observations are in contrast to that of competitive  $\text{Cd}^{2+}$  inhibition of  $\text{NO}_3^-$  uptake in a non- $\text{N}_2$ -fixing cyanobacterium (Singh and Yadava 1983), suggesting that the nature of  $\text{NO}_3^-$  – heavy-metal interaction may differ, not only from metal to metal, but also from organism to organism.

The apparent reduction of  $V_{\text{max}}$  for  $\text{NO}_3^-$  uptake (Fig. 3a) and NR activity rate (Fig. 4) with  $\text{Hg}^{2+} + \text{CH}_3\text{Hg}^+$  combinations over that of  $\text{Hg}^{2+}$ -treated *N. calcicola* cells, suggests that the inorganic and organic mercury species interact synergistically with each other. The comparatively lesser extent of reduction in  $V_{\text{max}}$  and activity rate with  $\text{Hg}^{2+} + \text{Cd}^{2+}$  combinations, suggests a lesser degree of synergism between these cations. Such synergistic interactions suggest that  $\text{Hg}^{2+}$  either becomes co-adsorbed on the common cell surface with  $\text{CH}_3\text{Hg}^+$  and  $\text{Cd}^{2+}$  or it facilitates the influx of such toxic cations, which result in the enhancement of inhibitory actions of  $\text{Hg}^{2+}$ , as reported in terms of photosynthetic  $\text{O}_2$  evolution and  $^{14}\text{CO}_2$  incorporation in *N. calcicola* (Singh and Singh 1987). On the other hand, the enhancement of  $V_{\text{max}}$  and enzyme activity rates with  $\text{Hg}^{2+} + \text{Ca}^{2+}$  or  $\text{Ni}^{2+}$  combinations over that of sets containing only  $\text{Hg}^{2+}$ , indicates antagonism between such cation combination (Figs. 3b, 4). The reduced degree of antagonism for  $\text{Hg}^{2+} + \text{Ni}^{2+}$  combinations over that of  $\text{Hg}^{2+} + \text{Ca}^{2+}$  may result either from direct competition for common uptake/binding site(s), or saturation of such site(s) preferentially by  $\text{Ni}^{2+}$  in the presence of  $\text{Hg}^{2+}$ , as reported for *Anabaena inaequalis* (Stratton and Corke 1979).  $\text{Cu}^{2+}$  ions showed a dual nature of interaction with  $\text{Hg}^{2+}$  depending on the parameters, i.e. antagonism for  $\text{NO}_3^-$  uptake and synergism for NR activity, which may be the result of its dual bonding preference for cellular ligands (Jones 1984).

**Acknowledgements.** We are grateful to Dr. S. Singh for helpful discussions and to the Head, Department of Botany and the Programme Coordinator, Centre of Advanced Study in Botany, for basic laboratory facilities. CBS gratefully acknowledges the Council of Scientific and Industrial Research (Government of India) for financial support in the form of a Research Associateship.

## References

- Allen MB, Arnon DI (1955) Studies on nitrogen-fixing blue-green algae. I. Growth and nitrogen fixation by *Anabaena cylindrica* Lemm. *Plant Physiol* 30:366–372
- Blinn DW, Tompkins T, Zaleski L (1977) Mercury inhibition on primary productivity using large volume plastic chambers in situ. *J Phycol* 13:58–61

- Flores E, Guerrero MG, Losada M (1983) Photosynthetic nature of nitrate uptake and reduction in the cyanobacterium *Anacystis nidulans*. *Biochim Biophys Acta* 722:408–416
- Guerrero MG, Lara C (1987) Assimilation of inorganic nitrogen. In Fay P, Van Baalen C (eds) *The Cyanobacteria*, Elsevier Science Publishers, pp 163–181
- Henriksen A, Wright RF (1978) Concentrations of heavy metals in small Norwegian lakes. *Water Res* 12:101–112
- Herbert D, Phipps PJ, Strange RE (1971) Chemical analysis of microbial cells. In Norris JR, Ribbons DW (eds) *Methods in microbiology*, vol VB, Academic Press, London, pp 209–344
- Jeffries TW (1982) The microbiology of mercury. *Progr Indust Microbiol* 76:21–75
- Jones MM (1984) Antagonists for toxic heavy metals. *Proc West Pharmacol Soc* 27:163–167
- Lowry OH, Rosenbrough NJ, Farr AL, Randall RJ (1951) Protein measurements with the Folin phenol reagent. *J Biol Chem* 193:265–275
- Mierle G, Stokes PM (1976) Heavy metal tolerance and metal accumulation by planktonic algae. In: Hemphill DD (ed) *Trace substances in environmental health*, vol XI, University Missouri, Columbia, pp 113–122
- Nicholas DJ, Nason A (1957) Determination of nitrate and nitrite. *Methods Enzymol* 3:981–984
- Rai LC, Dubey SK (1989) Impact of chromium and tin on nitrogen-fixing cyanobacterium *Anabaena doliolum*: interaction with bivalent cations. *Exotoxicol Environ Safety* 17:94–104
- Rai LC, Raizada M (1987) Toxicity of nickel and silver to *Nostoc muscorum*: interaction with ascorbic acid, glutathione and sulfur-containing amino acids. *Exotoxicol Environ Safety* 14:12–21
- Rai LC, Raizada M (1988) Impact of chromium and lead on *Nostoc muscorum*: regulation of toxicity by ascorbic acid, glutathione and sulfur-containing amino acids. *Exotoxicol Environ Safety* 15:195–205
- Rai LC, Gaur JP, Kumar HD (1981) Phycology and heavy metal pollution. *Biol Rev* 56:99–151
- Shieh YJ, Barber J (1973) Uptake of mercury by *Chlorella* and its effect on potassium regulation. *Planta* 109:49–60
- Sigmon CF, Kania HJ, Beyers RJ (1977) Reduction in biomass and diversity resulting from exposure to mercury in artificial streams. *J Fish Res Bd Can* 34:493–500
- Singh CB, Singh SP (1987) Effect of mercury on photosynthesis in *Nostoc calcicola*: role of ATP and interaction heavy metal ions. *J Plant Physiol* 129:49–58
- Singh SP, Yadava V (1983) Cadmium-induced inhibition of nitrate uptake in *Anacystis nidulans*: interaction with other divalent cations. *J Gen Appl Microbiol* 29:297–304
- Singh SP, Yadava V (1986) Cadmium tolerance in the cyanobacterium *Anacystis nidulans*. *Biol Zentralbl* 105:539–542
- Singh CB, Verma SK, Singh SP (1987) Impact of heavy metals on glutamine synthetase and nitrogenase activity in *Nostoc calcicola*. *J Gen Appl Microbiol* 33:87–91
- Singh DP, Khare P, Bisen PS (1989) Effect of  $\text{Ni}^{2+}$ ,  $\text{Hg}^{2+}$  and  $\text{Cu}^{2+}$  on growth, oxygen evolution and photosynthetic electron transport in *Cylindrospermum* PU 942. *J Plant Physiol* 134:406–412
- Snell FD, Snell CT (1949) Nitrate. In: *Colorimetric methods of analysis*, 3rd edn, vol II. Van Nostrand, New York, pp 785–807
- Starodub ME, Wong PTS, Mayfield CI, Chau YK (1987) Influence of complexation and pH on individual and combined heavy metal toxicity to a freshwater green alga. *Can J Fish Aquat Sci* 44:1173–1180
- Stratton GW, Corke CT (1979) The effect of mercuric, cadmium and nickel ion combinations on a blue-green alga. *Chemosphere* 10:731–740
- Stratton GW, Huber AL, Corke CT (1979) Effect of mercuric ion on the growth, photosynthesis and nitrogenase activity of *Anabaena inaequalis*. *Appl Environ Microbiol* 38:537–543
- Whitton BA (1970) Toxicity of heavy metals to algae: a review. *Phykos* 9:116–125